

Utilization of Peanut (*Arachis hypogaea*) Hull Based Activated Carbon for the Removal of Amaranth Dye from Aqueous Solutions

Zafar, Muhammad Nadeem*[†]; Ghafoor, Sabiha; Tabassum, Mehwish;
Zubair, Muhammad; Nazar, Muhammad Faizan; Ashfaq, Muhammad

Department of Chemistry, University of Gujrat, Gujrat, 50700 PAKISTAN

ABSTRACT: This research work is concerned with the investigation of removing amaranth foodstuff dye on activated carbon derived from peanut hulls (ACPH) as a low-cost adsorbent. The influence of different reaction parameters affecting dye uptake including pH, shaking speed, particle size, temperature, initial dye concentration, and contact time was investigated for proper selection of the optimized parameters for the removal process. By using ACPH, 76.92 mg/g adsorption capacity was achieved under agitation time of 90 min at pH 4 and temperature 60°C. Results demonstrate that experimental equilibrium data were fitted to Langmuir isotherm to a greater extent than Freundlich isotherm. Kinetics of amaranth dye on peanut hulls was found to follow pseudo-second-order kinetics. Results indicated that ACPH is a good adsorbent for removing amaranth dye from wastewater.

KEYWORDS: Amaranth; Peanut hull; Activated carbon; Kinetic modeling.

INTRODUCTION

Most industries, especially leather, paper, textile, food, cosmetic and plastic use dyes to color their products [1, 2]. The dyes (soluble or insoluble) are generally synthetic in nature, having intricate molecular structures and embodied with different functional groups that make them stable and tough to biodegrade [3]. Dyeing process in textile industries is one of the basic causes of contamination being participating in environmental pollution. Annually, around 10,000 different types of dyes mostly azo dyes and pigments are produced worldwide [4]. Dyes concentration in textile industries ranges from 10-10,000 ppm depended upon the dye removing treatments [5]. Wastewater discharged by

several industries e.g. textile industry, contains amaranth dye that is undesirable for environment due to its offensive side effects [2, 6-8]. A variety of products requires the use of amaranth dye in manufacturing process and can be found in the effluent of the majority sewage treatment services. There are a number of methods which have been developed to remove various health hazardous chemicals from waste material like organic and inorganic impurities [9]. These include: reverse osmosis [10, 11], biological treatment, advanced oxidation processes, coagulation [12, 13] and flocculation, photocatalytic process etc. [14]. Treatment by using these methods showed color reduction in dye solution but these are expensive method for decolonization.

* To whom correspondence should be addressed.

† E-mail: znadeempk@gmail.com ; nadeem.zafar@uog.edu.pk
1021-9986/2020/4/183-191 9/\$/5.09

A search for more effective method is important because of economical reasons and adsorption is considered to be an effective and economical strategy for dyes and other pollutants removal from wastewater [13, 15-23]. The chief adsorbent that is used in industry for adsorption purpose is the activated carbon prepared by using various sources [24]. Powdered peanut (*Arachis hypogaea*) hulls are used as activated carbon for the removal of dyes because nuts of agricultural wastes are available locally, economical [25, 26] and accessible from the 3500 acreage following yearly production of 140,000 tons of peanuts in Pakistan. Peanut hulls are generally available in markets as waste material so these are used to remove amaranth dye from wastewater. The present investigation reports a simple method for removal of amaranth dye by activated carbon derived from peanut hulls. For this purpose, various factors affecting the biosorption, such as pH of solution, shaking speed, particle size, temperature and contact time were investigated by the batch equilibration technique.

EXPERIMENTAL SECTION

Chemicals and instruments

Sodium hydroxide (NaOH), hydrochloric acid (HCl) and phosphoric acid (H₃PO₄) were purchased from Sigma (Sigma-Aldrich, Taufkirchen, Germany). Amaranth dye having molecular weight of 604.473 g/mol was purchased from Merck (Merck Damstadt, Germany) and its structure is given in Fig. 1.

A pH meter (inoLab pH 720, WTW, Weilheim, Germany) was employed for the pH measurements. The spectrophotometric measurements were carried out with a UV-Vis spectrophotometer (UV 4000, MRI, Germany). The solutions were stirred on orbital shaker (Wisd, WiseShake SHO-2D, Seoul, South Korea).

Preparation of adsorbent

Peanut (*Arachis hypogaea*) hulls were collected from local market of Kotla Arab Ali Khan (Pakistan) and dried in air. Fine powder of peanut hulls was prepared by grinding. In order to convert it into activated carbon, 10 g of this fine powder was taken and 20 ml of H₃PO₄ was added into it followed by stirring at 105°C. After that the mixture was heated in furnace at 405°C for 2 hours and after washing it was allowed to dry in oven overnight [27]. This activated carbon derived from peanut hulls (ACPH) was used as adsorbent for removal of amaranth dye.

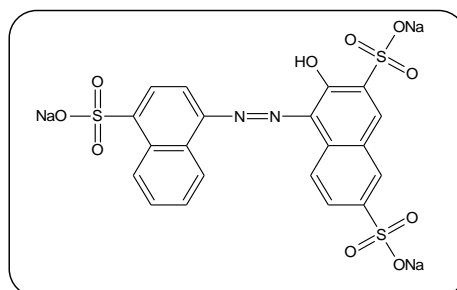


Fig. 1: Chemical structure of amaranth dye.

Equilibrium adsorption studies

Adsorption study was performed to find the effect of various factors on the adsorption behavior of amaranth dye onto ACPH. Different factors that control adsorption such as pH, stirring speed, particle size, temperature and contact time were optimized while studying the adsorption isotherm and kinetics of the reactions. Dye solutions with known concentrations were prepared and required amount of ACPH was added followed by shaking on orbital shaker for predefined time. Then the solutions were filtered and residual concentration of amaranth dye in the filtrate was determined by monitoring the absorbance at 521 nm using UV-Vis spectrophotometer. Adsorption capacity and percentage removal were calculated according to following equations 1 and 2 respectively [28].

$$q_e = (C_o - C_e) \times \frac{v}{w} \quad (1)$$

$$\% \text{ of dye removal} = \frac{C_o - C_e}{C_o} \times 100 \quad (2)$$

Where, C_o is the initial concentration (mg/L), C_e is the equilibrium concentration (mg/L) of dye, V is the volume (L) and w is the weight of adsorbent (g).

RESULTS AND DISCUSSION

Effect of different parameters on adsorption of amaranth dye on ACPH

The pH of the solution is found to have significant effect on the properties of the adsorbate and adsorbent. Influence of pH on the adsorption of amaranth dye was studied by varying the pH from 3-11 by keeping all the other factors constant. The results are shown in Fig. 2A.

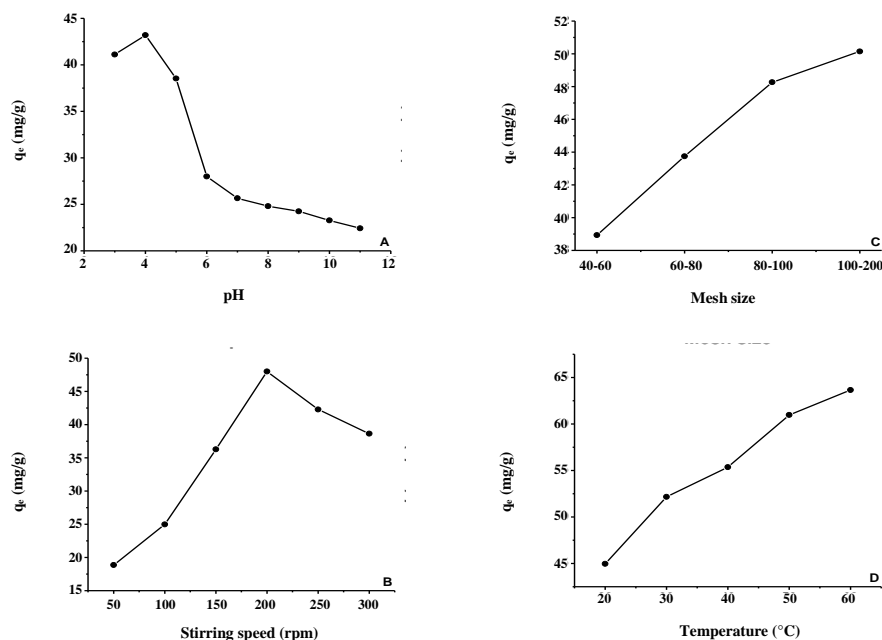


Fig. 2: Effect of the solution pH (A), stirring speed (B), adsorbent particle size (C), temperature (D) on the adsorption capacity of amaranth dye onto ACPH.

From the graph one can see that the maximum adsorption of dye was found at pH 4 and further increase in pH results in decrease in adsorption capacity. The point of zero charge (pH_{PZC}) of the adsorbent was found to be 6. The adsorbent potential surface became neutral, as pH of aqueous solution was equal to pH_{PZC} and is negatively charged when $\text{pH} > \text{pH}_{\text{PZC}}$ while it gets positively charged when $\text{pH} < \text{pH}_{\text{PZC}}$. The electrostatic attraction between the dye molecules (negatively charged) and ACPH surface (positively charged; $\text{pH}_{\text{PZC}} = 6$) might be the predominant adsorption mechanism at acidic pH. The increase in pH value caused a decrease in positive charge of the ACPH surface. This decreased the electrostatic forces among ACPH and amaranth dye, so there is reduction in adsorption of dye as the pH increased from 5 to 11.

The influence of stirring speed on adsorption process was studied by varying the stirring speed from 50 to 300 rpm at pH 4 and results are shown in Fig. 2B. The results showed that the adsorption capacity of amaranth dye

was increased upto 200 rpm, when shaking speed increased from 50 to 300 rpm. The possible explanation of this increase could be that as the shaking speed increased, the collision between dye and adsorbent surface was also increased which results in fast reaction between positively charged ACPH and anionic dye. At higher shaking speed a decrease in adsorption capacity of dye was observed.

Particle size of adsorbent affects the adsorption process, as smaller particle size would have greater proficiency to adsorb dye due to its larger surface area. The ratio of dye adsorbed approached the maximum value when the sorbent particle size was very small. Different mesh sizes from 40 to 200 microns were prepared from ACPH and influence of mesh size was studied at pH 4 and stirring speed 200 rpm. The results presented in Fig. 2C showed that the highest adsorption capacity was achieved by 100–200 mesh size. For convenience the mesh size 80–100 was used in further experiments.

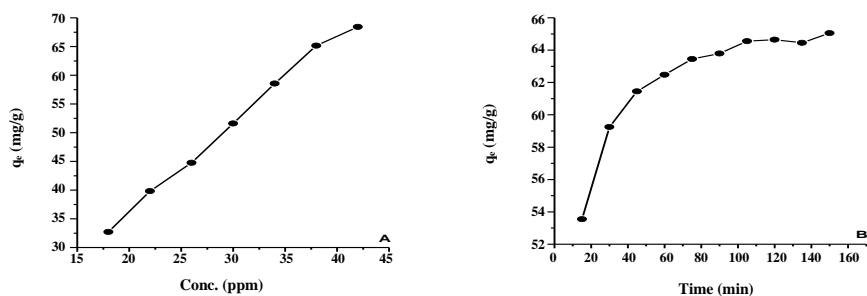


Fig. 3: Effect of initial dye concentration (A) and contact time (B) on the adsorption capacity of amaranth dye onto ACPH.

The effect of temperature was studied by varying the temperature from 20°C to 60°C at pH 4, stirring speed 200 rpm and mesh size 80–100. The obtained results (Fig. 2D) showed that the increase in the temperature of the solutions of amaranth dye from 20 to 60°C leads to an increase in the adsorption capacity of ACPH, which shows that the adsorption process is endothermic and chemical in nature. The possible explanation of this increase in adsorption capacity of amaranth dye onto ACPH could be due to the availability of more active sites and activation of the adsorbent surface at higher temperatures and increased diffusion and mobility of amaranth dye ions from the bulk solution towards the ACPH surface.

The apparent capacity of ACPH for amaranth dye was determined at the different concentrations. Fig. 3A clarifies the relation between capacities and the dye ion concentrations, which shows that as the dye ion concentration increased the adsorption capacity increased until 61.50 mg/g at 42 ppm. In general, the data indicated that sorption capacity increased with increase in initial dye ion concentration on ACPH. This sorption characteristic indicated that surface saturation was dependent on the initial dye ion concentrations. At low concentrations, adsorption sites took up the available dye more quickly. However, at higher concentrations, dye needed to diffuse to the biomass surface by intraparticle diffusion and greatly hydrolyzed ions will diffuse at a slower rate. The maximum dye ion sorption of an adsorbent may be determined from column experiments, by the use of a large excess of the adsorbate.

The contact time has a significant effect on adsorption process. The adsorption capacity of amaranth dye

was determined by varying the contact time from 15–150 min at pH 4, stirring speed 200 rpm, mesh size 80–100 and temperature 60°C. The results are shown in Fig. 3B. The rate of dye adsorption was rapid in the first 30 min and equilibrium was established in 80 min where adsorption efficiency reached its maximum value. After that no further increase in adsorption capacity was observed. At start more adsorption occurred due to availability of vacant sites on the surface of adsorbent but after that saturation occurred and adsorbate molecules could not enter in the inner sites due to high degree of saturation.

Equilibrium modeling

The equilibrium data of amaranth dye were analyzed by fitting into Langmuir and Freundlich equations to find out the suitable model that may be used for design consideration. Table 1 summarizes the constants and coefficients of different models. The linear form of Langmuir isotherm is given by the following equation [29]:

$$\frac{1}{q_e} = \frac{1}{q_m} + \frac{1}{K_L q_m} \times \frac{1}{C_e} \quad (3)$$

Where q_m is the maximum adsorption capacity relative to complete monolayer capacity (mg/g), K_L is Langmuir constant. Values of q_m and K_L are determined from the linear regression plot of $1/C_e$ and $1/q_e$ (Fig. 4A). Linear plot depicts the Langmuir adsorption plot revealing the preferential monolayer formation.

The characteristics of Langmuir isotherm can be expressed by a dimensionless constant, known as separation factor (R_L), which can be represented as:

$$R_L = \frac{1}{1 + K_L C_0} \quad (4)$$

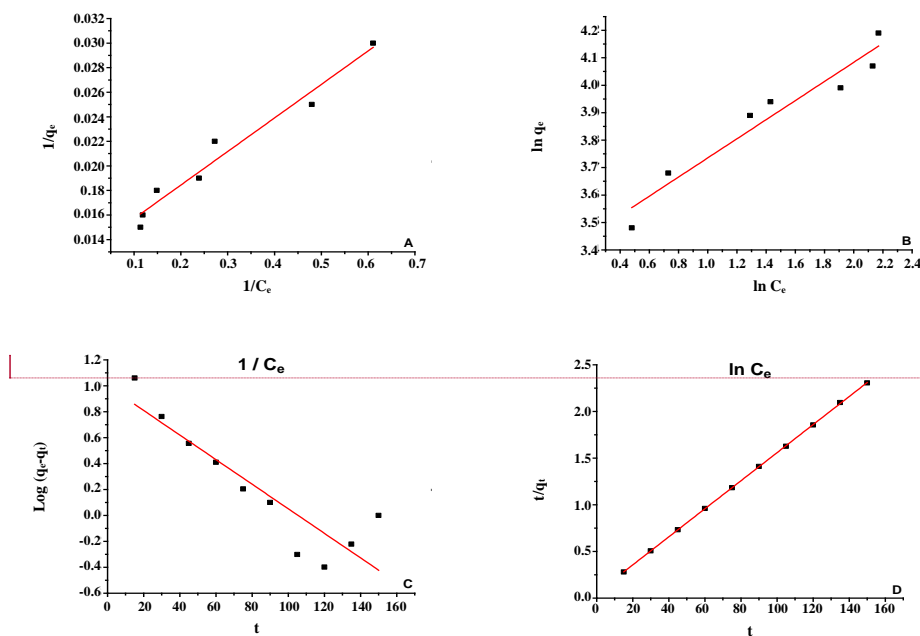


Fig. 4: Langmuir isotherm model (A), Freundlich isotherm model (B), Pseudo first order kinetic model (C) and Pseudo second order kinetic model (D) for amaranth dye removal using ACPH.

R_L value is an indication of adsorption nature to be unfavorable ($R_L > 1$), linear ($R_L = 1$), favorable ($0 < R_L < 1$) or irreversible ($R_L = 0$). R_L value for the adsorption of amaranth dye is found to be 0.047, which shows that adsorption is favorable.

The Freundlich isotherm involves multilayer formation on the surface of heterogeneous systems. The linear form of Freundlich equation is given as [30]:

$$\ln q_e = \ln K_F + \frac{1}{n} \ln C_e \quad (5)$$

Where K_F is the distribution coefficient tells the amount of dye adsorbed on adsorbent for unit equilibrium concentration, n is the adsorption intensity. Values of K_F and n are determined from the linear regression plot of $\ln q_e$ versus $\ln C_e$ (Fig. 4B). The value of n is found to be more than 1 and the value of $1/n$ is lie between ($0 < 1/n < 1$) so attributed to cooperative adsorption of dye and

revealed chemisorption process. The R^2 value of Freundlich model was lower than that obtained from Langmuir isotherm and also the calculated maximum adsorption capacity (q_m) from Langmuir isotherm agree well with experimental adsorption capacity showing that the equilibrium data best fitted to Langmuir Isotherm than Freundlich isotherm model which suggest monolayer adsorption of amaranth dye on the ACPH.

Kinetic modeling

Kinetics analysis for adsorption process can predict the rate at which a pollutant is removed from aqueous solutions and provides valuable data for understanding the mechanism of adsorption process. Pseudo first order and pseudo second order kinetic models are used to evaluate the mechanism involved in the process of adsorption. A linear form of pseudo first order equation is given as [31]:

Ce and ln Ce from /Please remove \ : [1] Commented [MNZ]
 .Figures 4c and 4d

Table 1: Adsorption isotherms parameters and adsorption kinetics constants of amaranth on ACPH.

Isotherms	Parameters	Values	
Langmuir	R_L	0.047	
	K_L (L/g)	0.481	
	q_m (mg/g)	76.92	
	R^2	0.956	
Experimental adsorption capacity	q_e (mg/g)	68.42	
Freundlich	K_F (mg/g)	29.55	
	n	2.870	
	R^2	0.922	
Pseudo first order	k_1 (1/min)	-0.022	
	q_e (mg/g)	9.980	
	R^2	0.784	
Experimental adsorption capacity	q_e (mg/g)	65.05	
Pseudo second order	k_2 (g/mg.min)	4.06×10^{-3}	
	q_e (mg/g)	66.53	
	R^2	0.999	
Thermodynamic	ΔH (kJ/mol)	18.50	
	ΔS (J/mol.K)	42.55	
	ΔG° (kJ/mol)	20°C	-2.450
		30°C	-4.870
		40°C	-6.120
		50°C	-8.340
60°C		-11.45	

$$\log(q_e - q_t) = \log q_e - \frac{k_1}{2.303} t \quad (6)$$

Where q_t is the amounts of adsorbed dye (mg/g) at time t and k_1 is the pseudo first order rate constant. Values of q_e and k_1 are calculated from the plot of $\log(q_e - q_t)$ versus t (Fig. 4C). The value of the correlation factor $R^2 = 0.784$, and the value of theoretical q_e is greatly differ from the experimental value suggesting that pseudo-first order model not explain the kinetics of adsorption (Table 1).

The linear form of pseudo second order model is given in Eq. (7) [32, 33]:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \quad (7)$$

Where k_2 is the pseudo second order rate constant. A plot between t and t/q_e was used to determine the values of q_e and k_2 (Fig. 4D). The value of correlation coefficient $R^2 = 0.999$ which is higher than that obtained from pseudo first order model. In addition, the theoretical

q_e value agrees with the experimental q_e value. The results obtained reflect that the adsorption of amaranth dye precedes second order kinetics mechanism. The values of constants and coefficients of different models are given in Table 1.

Thermodynamic modeling

Thermodynamic parameters such as, free energy change (ΔG°), enthalpy (ΔH) and entropy (ΔS) were evaluated to confirm the nature of adsorption of amaranth dye onto ACPH and can be estimated from the following equations:

$$\Delta G^\circ = -RT \ln K_d \quad (8)$$

$$\ln K_d = -\frac{\Delta H}{RT} + \frac{\Delta S}{R} \quad (9)$$

Where K_d is the thermodynamic equilibrium constant. The values of ΔH and ΔS were calculated from the slope and intercept by plotting $\ln K_d$ versus $1/T$ using

Table 2: Comparison of amaranth dye adsorption using ACPH with other reported adsorbents.

Adsorbent	Adsorption capacity (mg/g)	Reference
<i>Arachis hypogaea</i> shells	10.53	[34]
Powered peanut hull	14.90	[24]
Potato peels	1.710	[35]
Bottom ash	7.860	[36]
De-Oiled soya	12.70	[36]
MgAlCO ₃	120.9	[37]
Water hyacinth leaves	70.00	[38]
Core/Shell nanocomposites	142.1	[39]
ACPH	76.92	This study

equation 9 and the values of ΔG° were calculated according to equation 8.

Table 1 summarizes the values of ΔG° , ΔH and ΔS . The positive values of ΔH and ΔS show that the adsorption process is endothermic with increasing the randomness of the system and dominant by chemisorption. The negative value of free energy indicates that the adsorption process is spontaneous.

Table 2 showed the comparison of amaranth dye's theoretical adsorption capacity onto ACPH with other reported adsorbents. It can be perceived that the ACPH has a comparable and good adsorption capacity, which showed that ACPH could be suitable adsorbent for azo dyes removal.

CONCLUSIONS

Activated carbon derived from waste material powdered peanut hulls were effectively utilized, as adsorbent for the removal of amaranth dye from aqueous solution, since the peanut hulls were easily available in large quantity in Pakistan and the treatment method seemed to be economical. The maximum adsorption capacity of amaranth dye onto ACPH was achieved at initial pH 4, stirring speed 200 rpm, mesh size 80-100 and temperature 60°C. Isothermal data of adsorption process followed Langmuir model. The kinetics study clearly demonstrates that adsorption follow pseudo-second-order kinetics. The adsorption capacity was found to be 76.92 mg dye per gram of the adsorbent for amaranth. Thus adsorption can be a feasible for wastewater treatment and ACPH can act as a good and economical adsorbent.

Acknowledgements

The authors are thankful to Chemistry Department, University of Gujrat, Pakistan for providing research facilities to conduct this research. We have no conflicts of interest to disclose.

Received : Nov. 23, 2018 ; Accepted :Apr. 15, 2019

REFERENCES

- [1] Yesilada O., Asma D., Cing S., [Decolorization of Textile Dyes by Fungal Pellets](#), *Process Biochem.*, **38**(6): 933-938 (2003).
- [2] Gottlieb A., Shaw C., Smith A., Wheatley A., Forsythe S., [The Toxicity of Textile Reactive Azo Dyes After Hydrolysis and Decolourisation](#), *J. Biotechnol.*, **101**(1): 49-56 (2003).
- [3] Sulyman M., Namiesnik J., Gierak A., [Utilization of New Activated Carbon Derived from Oak Leaves for Removal of Crystal Violet from Aqueous Solution](#), *Pol. J. Environ. Stud.*, **23**(6): 2223-2232 (2014).
- [4] Rafii F., Hall J., Cerniglia C., [Mutagenicity of Azo Dyes Used in Foods, Drugs and Cosmetics before and After Reduction by Clostridium Species from the Human Intestinal Tract](#), *Food Chem. Toxicol.*, **35**(9): 897-901 (1997).
- [5] Ho, Y.S., McKay, G., [Sorption of Dye from Aqueous Solution by Peat](#), *Chem. Eng. J.*, **70**(2): 115-124 (1998).
- [6] Ganesh R., Boardman G.D., Michelsen D., [Fate of Azo Dyes in Sludges](#), *Water Res.*, **28**(6): 1367-1376 (1994).

- [7] Wang C., Yediler A., Lienert D., Wang Z., Ketrup A., Toxicity Evaluation of Reactive Dyestuffs, Auxiliaries and Selected Effluents in Textile Finishing Industry to Luminescent Bacteria *Vibrio Fischeri*, *Chemosphere*, **46**(2): 339-344 (2002).
- [8] Pearce C., Lloyd J., Guthrie J., The removal of Colour from Textile Wastewater Using Whole Bacterial Cells: A Review, *Dyes Pigm.*, **58**(3): 179-196 (2003).
- [9] Poul M., Jarry G., Elhkim M.O., Poul J.M., Lack of Genotoxic Effect of Food Dyes Amaranth, Sunset Yellow and Tartrazine and Their Metabolites in the Gut Micronucleus Assay in Mice, *Food Chem. Toxicol.*, **47**(2): 443-448 (2009).
- [10] Al-Bastaki N., Removal of Methyl Orange Dye and Na₂SO₄ Salt from Synthetic Waste Water Using Reverse Osmosis, *Chem. Eng. Process-Process Intensification*, **43**(12): 1561-1567 (2004).
- [11] Suksaroj, C., Heran, M., Allegre, C., and Persin, F., Treatment of Textile Plant Effluent by Nanofiltration and/or Reverse Osmosis for Water Reuse, *Desalination*, **178**(1): 333-341 (2005).
- [12] Aleboyeh, A., Daneshvar, N., and Kasiri, M.B., Optimization of CI Acid Red 14 Azo Dye Removal by Electrocoagulation Batch Process with Response Surface Methodology, *Chem. Eng. Process-Process Intensification*, **47**(5): 827-832 (2008).
- [13] Beltrán-Heredía J., Sánchez Martín J., Azo Dye Removal by Moringa Oleifera Seed Extract Coagulation, *Coloration Technol.*, **124**(5): 310-317 (2008).
- [14] Gupta V.K., Ali I., Saleh T.A., Nayak A., Agarwal S., Chemical Treatment Technologies for Waste-Water Recycling—an Overview, *RSC Adv.*, **2**(16): 6380-6388 (2012).
- [15] Acar E.T., Ortaboy S., Atun G., Adsorptive Removal of Thiazine Dyes from Aqueous Solutions by oil Shale and Its Oil Processing Residues: Characterization, Equilibrium, Kinetics and Modeling Studies, *Chem. Eng. J.*, **276**: 340-348 (2015).
- [16] Liao P., Malik Ismael Z., Zhang W., Yuan S., Tong M., Wang K., Bao J., Adsorption of Dyes from Aqueous Solutions by Microwave Modified Bamboo Charcoal, *Chem. Eng. J.*, **195-196**: 339-346 (2012).
- [17] Pietrelli L., Francolini I., Piozzi A., Dyes Adsorption from Aqueous Solutions by Chitosan, *Sep. Sci. Technol.*, **50**(8): 1101-1107 (2015).
- [18] Rani S., Sumanjit K., Mahajan R.K., Comparative Study of Surface Modified Carbonized Eichhornia crassipes for Adsorption of Dye Safranin, *Sep. Sci. Technol.*, **50**(16): 2436-2447 (2015).
- [19] Rahman, A., Kishimoto, N., and Urabe, T., Adsorption Characteristics of Clay Adsorbents – Sepiolite, Kaolin and Synthetic Talc – for Removal of Reactive Yellow 138:1, *Water Environ. J.*, **29**(3): 375-382 (2015).
- [20] Zafar M.N., Nadeem R., Hanif M.A., Biosorption of Nickel from Protonated Rice Bran, *J. Hazard. Mater.*, **143**(1-2): 478-485 (2007).
- [21] Haq A.u., Shah J., Jan M.R., Din S.u., Kinetic, Equilibrium and Thermodynamic Studies for the Sorption of Metribuzin from Aqueous Solution Using Banana Peels, an Agro-Based Biomass, *Toxicol. Environ. Chem.*, **97**(2): 124-134 (2015).
- [22] Ishaq M., Saeed K., Ahmad I., Sultan S., Akhtar S., Coal Ash as a Low-Cost Adsorbent for the Removal of Xylenol Orange from Aqueous Solution, *Iran. J. Chem. Chem. Eng. (IJCCCE)*, **33**(1): 53-58 (2014).
- [23] Bouguettoucha A., Reffas A., Chebli D., Amrane A., Adsorption of the Cationic Dye Ethyl Violet on Acid and Alkali-Treated Wild Carob Powder, A Low-Cost Adsorbent Derived from Forest Waste, *Iran. J. Chem. Chem. Eng. (IJCCCE)*, **36**(1): 87-96 (2017).
- [24] Gong R., Ding Y., Li M., Yang C., Liu H., Sun Y., Utilization of Powdered Peanut Hull as Biosorbent for Removal of Anionic Dyes from Aqueous Solution, *Dyes Pigm.*, **64**(3): 187-192 (2005).
- [25] Raffa R.B., Wu C., Stone D.J., Borenstein M.R., Codd E.E., Coogan T.P., Determination of the Adsorption of Tramadol Hydrochloride by Activated Charcoal *in Vitro* and *in Vivo*, *J. Pharmacol. Toxicol. Methods*, **43**(3): 205-210 (2000).
- [26] Gong R., Sun Y., Chen J., Liu H., Yang C., Effect of Chemical Modification on Dye Adsorption Capacity of Peanut Hull, *Dyes Pigm.*, **67**(3): 175-181 (2005).

- [27] Yang C., Ke L.X., Gong R.M., Liu H.J., Sun Y.Z., Utilization of Powdered Peanut Hull as Biosorbent for Removal of Azo Dyes from Aqueous Solution, *J. Biol.*, **2**: 016 (2005).
- [28] Chang I.-S., Le Clech P., Jefferson B., Judd S., Membrane Fouling in Membrane Bioreactors for Wastewater Treatment, *J. Environ. Eng.*, **128**(11): 1018-1029 (2002).
- [29] Langmuir, I., Adsorption of Gases on Plane Surfaces of Glass, Mica and Platinum, *J. Am. Chem. Soc.*, **40**: 1361-1403 (1918).
- [30] Freundlich, H., Adsorption in Solutions, *J. Phys. Chem.*, **57**: 384-410 (1906).
- [31] Lagergren, S., About the Theory of So-Called Adsorption of Soluble Substances, *K. Sven. Vetensk.akad. Handl.*, **24**(4): 1-39 (1898).
- [32] Ho Y.S., McKay G., Pseudo-Second order Model for Sorption Processes, *Process Biochem.*, **34**(5): 451-465 (1999).
- [33] Zafar M.N., Aslam I., Nadeem R., Munir S., Rana U.A., Khan S.U.-D., Characterization of Chemically Modified Biosorbents from Rice Bran for Biosorption of Ni(II), *J. Taiwan Inst. Chem. Eng.*, **46**: 82-88 (2015).
- [34] Rehman R., Afzal A., Batch Scale Removal of an Organic Pollutant Amaranth Dye from Aqueous Solution Using *Pisum sativum* Peels and *Arachis hypogaea* Shells as Adsorbents, *J. Chem. Soc. Pak.*, **37**: 930-938 (2015).
- [35] Rehman, R., Mahmud T., Irum M., Comparative Sorption Studies for Amaranth Dye Removal from Water in a Cost-Effective Way Using Guava Leaves and Potato Peels, *Asian J. Chem.*, **27**(6): 2008-2014 (2015).
- [36] Mittal A., Kurup L., Gupta V.K., Use of Waste Materials—Bottom Ash and De-Oiled Soya, as Potential Adsorbents for the Removal of Amaranth from Aqueous Solutions, *J. Hazard. Mater.*, **117**(2): 171-178 (2005).
- [37] Abdellaoui K., Pavlovic I., Bouhent M., Benhamou A., Barriga C., A Comparative Study of the Amaranth Azo Dye Adsorption/desorption from Aqueous Solutions by Layered Double Hydroxides, *Appl. Clay Sci.*, **143**: 142-150 (2017).
- [38] Guerrero-Coronilla I., Morales-Barrera L., Cristiani-Urbina E., Kinetic, Isotherm and thermodynamic Studies of Amaranth Dye Biosorption from Aqueous Solution onto Water Hyacinth Leaves, *J. Environ. Manage.*, **152**: 99-108 (2015).
- [39] Huo Y., Wu H., Wang Z., Wang F., Liu Y., Feng Y., Zhao Y., Preparation of Core/Shell Nanocomposite Adsorbents Based on Amine Polymer-Modified Magnetic Materials for the Efficient Adsorption of Anionic Dyes, *Colloids Surf. A: Physicochem. Eng. Asp.*, **549**: 174-183 (2018).